

Appendix L

System Assessment Capability: A 10,000-Year, Post-Closure Assessment

L.1 Introduction

In late 1997, the U.S. Department of Energy (DOE) established the Groundwater/Vadose Zone Integration Project with Bechtel Hanford, Inc. (BHI), the Hanford Site Environmental Restoration Contractor, as manager. The project transitioned to Fluor Hanford, the Project Hanford Management Contractor, in July 2002, and has been renamed the Groundwater Protection Program. Pacific Northwest National Laboratory (PNNL) is a partner in the project. The mission of the project is to coordinate and integrate projects that characterize, monitor, and clean up contaminants in the groundwater and vadose zone (the soil between the ground surface and the groundwater) beneath the Hanford Site. The Groundwater Protection Project also incorporates other task areas that complement these projects and several that represent accelerated actions leading to earlier site cleanup and closure.

In 1999, under the Integration Project, DOE initiated development of an assessment tool that will enable users to model the movement of contaminants from all waste sites at Hanford through the vadose zone, the groundwater, and the Columbia River and to estimate the impact of contaminants on human health, ecology and the local cultures and economy. This tool is named the System Assessment Capability (SAC).

The approach taken by the SAC is consistent with the methods, characteristics, and controls associated with a composite analysis as described by the Columbia River Comprehensive Impact Assessment (CRCIA) team (DOE-RL 1998). The CRCIA was a study initiated by DOE, the Washington State Department of Ecology, and the U.S. Environmental Protection Agency (EPA) to assess the effects of Hanford-derived materials and contaminants on the Columbia River environment, river-dependent life, and users of river resources. Part I of CRCIA was a study of present-day impacts to the Columbia River from Hanford contaminants. Part II was a suite of requirements for the development of a comprehensive impact assessment for the Columbia River. The two key elements of the SAC approach are 1) ensuring that factors that will dominate the risk are included, and 2) providing an understanding of the uncertainty of the results. Dominant factors were identified through scoping studies and the development of conceptual models for each of the analysis modules used. A stochastic modeling approach was taken to estimate uncertainty in the results. Aspects of uncertainty that could not be included in the calculation were considered in the analysis of the modeling results and discussed in the document presenting initial assessment results (Bryce et al. 2002). The analysis modules included in the SAC parallel those identified by CRCIA and were developed through work group meetings that included regulator and stakeholder participation.

1 Several key modules were adopted directly from the CRCIA, including the module used to calculate
2 human health impacts (the HUMAN code) and the module used to calculate impacts to ecological species
3 (the ECEM code).
4

5 An initial assessment was recently completed with the SAC to demonstrate its functional assessment
6 capability. Future modifications to the tool will be driven by the requirements of specific assessments.
7 Improvements in the results obtained from use of the SAC will be realized as input data are refined
8 through characterization and scientific research. Bryce et al. (2002) reported the results of that assess-
9 ment, which is the basis for application of the SAC to provide a site-wide perspective of waste disposal
10 and remedial actions in this Hanford Solid Waste Environmental Impact Statement (HSW EIS). Much of
11 the material presented in this appendix has been taken from Bryce et al. (2002).
12

13 To simplify the discussion presented in this appendix, the term “SAC” refers to the software package
14 used for this assessment, but it should be noted that the SAC is an evolving and maturing capability.
15

16 The initial assessment in fiscal year 2002:

- 17
- 18 • Modeled the movement of contaminants from 533 locations throughout the Hanford Site representing
19 890 waste sites through the vadose zone, the groundwater, and the Columbia River.
20
- 21 • Incorporated data on 10 radioactive and chemical contaminants—carbon tetrachloride, cesium-137,
22 chromium, iodine-129, plutonium-239/240, tritium, strontium-90, technetium-99, total uranium
23 (chemical), and uranium (radionuclide).
24
- 25 • Focused on subsurface transport, the Columbia River, and risks to human and ecological health, and
26 the economy and culture.
27
- 28 • Included the geographic region from Rattlesnake Mountain to the Columbia River and from Vernita
29 Bridge to McNary Dam on the Columbia River.
30
- 31 • Included the cleanup actions in Hanford’s cleanup plans and agreements as of October 2000.
32
- 33 • Consisted of a stochastic simulation for the period 1944 to 3050 using 25 realizations, thus providing
34 insight into the median response and an initial look at uncertainty.
35
- 36 • Simulated a 1000-year, post-closure period. Three waste forms known to release after that time were
37 not included—immobilized low-activity waste (ILAW), melters, and naval reactor compartments.
38

39 For the waste sites located on the Hanford Central Plateau and their associated contaminant plumes,
40 the findings of the initial assessment parallel those of the composite analysis (Kincaid et al. 1998). The
41 results are also consistent with concentrations in environmental media measured by the Hanford Envi-
42 ronmental Surveillance Program (Poston et al. 2002). Both the monitoring results and the assessment
43 reported here indicate that Hanford impacts to the Columbia River have peaked and are now declining.
44

1 For the purposes of the HSW EIS, the System Assessment Capability (SAC) is a ‘best available
2 technology’ and, while it remains a tool under development, the SAC Rev. 0 tool is adequate to provide
3 valuable information through quantification of cumulative risks and impacts associated with solid waste
4 disposal at the Hanford Site.

6 **L.1.1 Context of SAC Runs**

8 The principal SAC simulation made in support of the HSW EIS is a series of 25 stochastic simula-
9 tions run over the period 1944 through 12050 A.D. (that is, a 10,000-year, post-closure period), for the
10 Hanford Site Disposition Baseline (HSDB) scenario. This simulation includes a stochastic representation
11 of inventory, release and transport, and a deterministic representation of exposure and dose. In addition, a
12 median-value input case, based on the median value of each input parameter represented by a distribution
13 in the stochastic model, was simulated.

15 The HSDB scenario represented in the fiscal year 2002 initial assessment are based on a number of
16 cleanup assumption including waste, debris, and contaminated soil will be removed from the 100 Areas
17 and the remaining soil will meet residential use standards. Similarly, waste, debris, and contaminated soil
18 will be removed from the 300 Areas, but the remaining soil will meet industrial use standards. In this
19 scenario, retrievably stored transuranic (TRU) waste will be recovered, tested to determine waste content,
20 repackaged, and sent offsite for disposal at the Waste Isolation Plant in New Mexico. The waste in Burial
21 Grounds 618-10 and 618-11 will be removed, and the TRU waste will be repackaged and removed from
22 the Hanford Site, while the low-level waste (LLW) will be disposed of in solid waste disposal facilities in
23 the Central Plateau. Ninety-nine percent of the tank waste volume will be recovered from the tanks and a
24 1 percent residual volume will remain. Losses to the subsurface during waste recovery are assumed to
25 average 30,280 L (8000 gal) per single-shell tank recovered. The recovered tank waste will be separated
26 into low-activity and high-activity fractions. Both waste fractions are assumed to be immobilized. Low-
27 activity waste will be disposed of onsite, while the high-activity fraction will be disposed of in the
28 national repository. All spent fuel also will be stored in a stable configuration for shipment to and
29 disposal in the national repository.

31 The initial assessment and this analysis assume that, for the duration of the analysis, the future
32 regional and local climate will remain unchanged for the period of the analysis. Furthermore, it is
33 assumed that major engineered structures in the region (for example, the reservoir system on the
34 Columbia River) will remain in place. The recorded climate and environmental response (for example,
35 Columbia River stage and discharge records) since startup of the site operations were used to simulate the
36 period from 1944 to the present. The climate record from 1961 to 1990 was used to represent the future
37 climate. Consequently, the Hanford Site remains a semi-arid, shrub-steppe environment in the simula-
38 tions. The riparian zone, Columbia River, and river ecosystem are assumed to remain essentially
39 unchanged for the duration of the analysis. Also, the human population will be unchanged and will be
40 based on the current socio-economic setting. Analyses of alternate future climates (for example, global
41 climate change or onset of an ice age and glacial flooding) and potential future events (for example, fail-
42 ure or removal of the reservoir system) are not addressed.

1 Where the initial assessment addressed the period 1944 through 3050 (that is, essentially a 1000-year,
2 post-closure simulation), simulations for this EIS were carried out over a 10,000-year, post-closure
3 period. Within the SAC, a single transport pathway element, the Columbia River model, is limited to the
4 year 10,000 A.D. in its simulation algorithm, but all other transport pathways (release, vadose zone,
5 groundwater) can execute for the full 10,000-year, post-closure period.
6

7 The stochastic simulations supporting the HSW EIS are based on the parameter distributions assem-
8 bled for the initial assessment. In addition to the environmental pathway and risk/impact model parame-
9 ters, the inventory and the future disposal and remedial actions assembled for the initial assessment are
10 included. Differences between the inventory used in this extended simulation of the initial assessment
11 and that used in the HSW EIS are described in Section L.2.2.2. Principal differences lie in the methods
12 used to forecast solid waste disposal actions until site closure, both for onsite generators (for example,
13 Waste Treatment Plant contributions) and for offsite generators.
14

15 The potential contaminants of greatest concern include technetium-99, iodine-129 and uranium.
16 These contaminants appear in solid waste performance assessments (Wood et al. 1995, Wood 1996) that
17 analyze solid waste disposals in 200 West and 200 East Areas. While the initial application of SAC to the
18 HSW EIS did not include iodine-129, an ability to achieve simulation of iodine-129 is being established.
19 Of necessity, simulation of iodine-129 will include an initial condition for iodine-129 representative of
20 prior releases to the unconfined groundwater, simulation of future releases of iodine-129 per the initial
21 assessment, and superposition of the ILAW contribution to iodine-129 risk and impact. This approach to
22 iodine-129 simulation will include events attributed to past liquid discharges (current groundwater
23 plumes), future solid waste releases, and long-term future releases from immobilized low-activity tank
24 waste. The inventory estimated to exist in the unconfined aquifer, and the estimate of iodine-129 in low-
25 activity tank waste to remain at Hanford will be used in this estimate of the iodine-129 contribution to
26 risk/impact. As in the original 1000-yr initial assessment, simulation of technetium-99 and uranium will
27 use the complete history and forecast of their disposal and begin in 1944 with a clean subsurface
28 environment.
29

30 It is unlikely that the plumes from these three classes of release events will superimpose in time. The
31 liquid discharge and unplanned release (e.g., tank leak) sites have created groundwater plumes and will
32 likely continue to release to groundwater during the immediate future. Releases from dry solid waste dis-
33 posals have some containment (e.g., boxes, drums, plastic bags) and less driving force (e.g., infiltration),
34 and, therefore, they will likely release later than the liquid releases. Finally, the substantially stable and
35 long-term waste forms like vitrified low-activity tank waste will not corrode and release for thousands of
36 years. It is unlikely that peaks from each of these types of release will superimpose in space and time.
37

38 **L.1.2 Relationship to EIS Calculations**

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40 The EIS calculations focus on the impacts associated with alternatives to the disposal of solid waste.
41 The SAC represents a holistic examination of the radioactive and chemical waste legacy of the Hanford
42 Site. For this reason, it can be used to examine the relative risk and impact associated with disposal and
43 remedial action alternatives and the relative role of different segments of Hanford waste—for example,
44 solid waste, past-practice liquid discharges, or tank wastes. Used in this way, the SAC provides an ability

1 to visualize the change in impact associated with various options and wastes. This kind of cumulative
2 impact assessment provides a larger scale site-wide context from which to view the alternatives and
3 influence disposal decisions.
4

5 The EIS calculations provide a detailed evaluation of each specific alternative. The SAC is only able,
6 at this time, to present the single case of an extended analysis (e.g., 10,000 yr post closure) of the HSDB.
7 In essence, the SAC provides an estimate of the contribution made to risk and impact from technetium-99
8 and uranium from other Hanford waste disposal and remedial actions not explicitly considered in the
9 HSW EIS alternatives, and to contrast that with the contribution from solid wastes.
10

11 **L.2 Methods and Approach**

12

13 Historically, DOE has used various tools to assess the effects of waste management and cleanup
14 activities on the environment. Assessments have been performed to address a range of questions. Some
15 assessments have focused on individual waste sites or waste types—for example the assessment per-
16 formed to evaluate the future performance of the glass waste form proposed for isolating low-activity
17 waste currently in tanks (Mann et al. 2001). Others have looked at contaminants from a variety of
18 sources. The Hanford Environmental Dose Reconstruction Project estimated human health impacts from
19 past releases to the atmosphere and river (Farris et al. 1994) during Hanford operations from 1944 to
20 1972. The Columbia River Comprehensive Impact Assessment (CRCIA) (DOE-RL 1998) examined
21 ecological and human health effects that might result from the 1990 to 1996 distribution of contaminants
22 in the environment in and near the Columbia River. The composite analysis performed in 1997 consid-
23 ered the impact of selected radionuclides from approximately 280 waste sites in the 200 Areas
24 (Kincaid et al. 1998). In 2001, Bergeron et al. (2001) issued an addendum to the composite analysis that
25 considered additional waste sites on the Central Plateau.
26

27 The collective impact of all of the waste that will remain at Hanford, however, had not yet been inte-
28 grated to provide an understanding of the cumulative effects of Hanford activities on the Central Plateau
29 as well as in the river corridor. The SAC was developed to fill this gap and has benefited from the lessons
30 learned in previous assessments.
31

32 The initial assessment and this extension to a 10,000-year, post-closure analysis considers solid waste
33 disposals in the Central Plateau as occurring within aggregated solid waste disposal facilities in the north-
34 ern and southern portions of the 200 West and East Areas. Annual inventories for each disposal facility
35 within a subregion of the site are aggregated to create an annual solid waste inventory for the subregion.
36 The areal footprints of disposal facilities within a subregion are aggregated to create a total solid waste
37 disposal facility areal footprint. Contaminants from the aggregated disposal facility are released to the
38 unconfined aquifer at the centroid coordinates of the aggregated disposal facility. Thus, use of an aggre-
39 gated representation of solid waste disposal facilities is an approximation in a number of ways. Notably,
40 the inventory actually placed in individual trenches within each disposal facility is represented as distrib-
41 uted over the entire areal footprint of the disposal facility. Hence, the aggregated inventory is distributed
42 over the aggregated areal footprint of all solid waste disposal facilities in a subregion of the site. Because